

¹¹² A typical wildland basin wildrye (Leymus cinereus) population.

Does basin wildrye (*Leymus cinereus*) show local adaptation when deployed according to generalized provisional seed zones in the Central Basin and Range ecoregion?

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ABSTRACT

Identifying genetically appropriate plant materials for seed-based restoration relies on the principle of local adaptation: the intent to match adaptive genetic characteristics to variation in ecological clines pertinent to plant establishment and persistence. Seed zone maps delineate some of these relationships. Generalized provisional seed zones (GPSZ) were developed for use where species-specific seed zones are lacking, as was the case for basin wildrye at the initiation of this study. In this study, basin wildrye (Leymus cinereus (Scribn. & Merr.) Á. Löve [Poaceae]) sources from 25 wild populations and 4 commercial varieties were planted at 4 test sites representing the species distribution across GPSZ in the central basin and range ecoregion. Sources were seeded separately into 5 blocks for each of 2 treatments and data were grouped by tetraploid or octoploid cytotype and local or non-local origin for comparison. Treatments included coverage with row cover or uncovered controls. We assessed initial and short-term survival. Plantings failed at 2 sites in both 2013 and 2014, with too few plants to quantify differences. At the remaining 2 sites we found no differences in initial survival or short-term persistence between pooled local sources compared to pooled non-local sources. Among commercial sources, the cultivar 'Magnar' initially outperformed local pooled materials at Fountain Green but not at Nephi. This difference was not evident 3 and 4 y later. Initial establishment under row cover was dramatically better than uncovered controls but persisted only through years 3 and 4 at the Fountain Green site.

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KEY WORDS

local adaptation, basin wildrye, *Leymus cinereus*, Great Basin, provisional seed zones, row cover, Poaceae

NOMENCLATURE

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n recent years, plant material programs and policies have emphasized the development and deployment of genetically appropriate native plant materials (Richards and Chambers 1996; Rogers and Montalvo 2004; Shaw and others 2008; Oldfield and Olwell 2015; Wood and others 2015). The foundation of genetic suitability is the principle of local adaptation with the intent of exploiting genetic variation to both preserve and capitalize on spatially diverse functional traits to improve restoration outcomes (Leimu and Fischer 2008). In application, using genetically appropriate plant materials requires a suitable pairing of restoration sites with seed sources, so that the developing plant community provides desired ecosystem services and persists through environmental challenges (Johnson and others 2010b; Jones 2013).

Genecological studies are a common first step in understanding morphological and phenological variation that can be used to identify adaptive genetic differences between populations (Campbell 1986; Erickson and others 2004; Johnson and others 2015). In genecological studies, numerous populations of the same species are planted in one or more common environments, and many traits that may have adaptive significance are evaluated. Resulting data, when paired with climate records from source population locations, are used to develop seed transfer guidelines, which are often displayed as seed zone maps. For a number of prominent Great Basin restoration grasses, genecological work is underway (Johnson and others 2010a; Johnson and others 2012; St Clair and others 2013; Johnson and others 2015) and was recently completed for basin wildrye (Leymus cinereus (Scribn. & Merr.) Á. Löve [Poaceae]) (Johnson and Vance-Borland 2016). Yet, where genecological work is absent, many species will, of necessity, be planted without species-specific seed zone recommendations (Bower and others 2014). In these cases, surrogate approaches have been suggested (Omernik 1987; Milberg and Andersson 1998; Hargrove and Hoffman 2005; Vogel and others 2005; Johnson and others 2010b; Miller and others 2011; USDA 2013).

In this study, we evaluated a leading surrogate, generalized provisional seed zones (GPSZ) (Bower and others 2014; Kramer and others 2015), as a method of matching basin wildrye source populations to representative restoration sites. GPSZ delineate areas of similar winter minimum temperature and annual aridity, characteristics important to plant adaptation. When paired with ecoregions (Omernik 1987), they partition areas into climatically and ecologically distinct units and are recommended as a starting point for developing seed transfer guidelines.

Basin wildrye was selected as the test species because it is a common restoration species (Krabacker 2013) in the Intermountain region, yet its genecological seed zones had not been developed at the time this study was initiated. Its distribution extends from New Mexico to Saskatchewan and then west to the Pacific Coast. Within the Central Basin and

Range (CBR) ecoregion, its known distribution ranges over 1100 m (3630 ft) in elevation and across 9, level 4 ecoregions (Omernik 1987). A substantial array of climatic and ecological variation occurs throughout this species distribution that may lead to the development of localized adaptions between spatially or ecologically divergent populations (Hereford 2009). Across the western portion of the species distribution, Culumber and others (2013) genetically distinguished 3 metapopulation races corresponding to the Columbia, Rocky Mountain, and Great Basin regions. Octoploid cytotypes were more abundant in the western portion of the species range, and tetraploid cytotypes were more abundant in the eastern portion. In the Great Basin, where the Great Basin race is fully encompassed by the more broadly distributed Rocky Mountain race, both cytotypes are common. In a recently completed common-garden study (Johnson and Vance-Borland 2016), genetic variation for both cytotypes was linked to source climates. That information was used to delineate 15 genecological seed zones for basin wildrye.

To test the principle of local adaptation, we hypothesized that wild basin wildrye sources planted into test sites matching their home GPSZ would demonstrate local adaptation through higher initial establishment and better short-term persistence. We also hypothesized local GPSZ pooled sources would similarly outperform the cultivars 'Magnar', 'Trailhead', and 'Continental', which originate from outside the Great Basin, and Great Basin 'Tetra', which is a composite of 31 sources from multiple GPSZ's and ecoregions. From experience using fabric row cover to improve seedling establishment in nursery beds, we included row cover as a treatment option in this experiment and hypothesized basin wildrye establishment would be better under row cover.

METHODS

Within the Great Basin, known populations of basin wildrye (n = 107) were mapped to 9 of the 20 Great Basin GPSZ's (WWETAC 2021) using ArcMap 10.1 (ESRI 2012). Five zones were poorly represented, containing 1 to 3 populations within their boundaries, while 4 zones accounted for the remaining 91.6% of populations. Bower and others (2014) demonstrated GPSZ's to be most effective when nested within Omernik's level III ecoregions. Consequently, for this study, source populations were restricted to origins within the CBR ecoregion. Basin wildrye population locations were obtained from a US Forest Service database developed over the course of several decades of plant material work and were originally located by travel throughout the area, consultation with federal and state agency personnel, or university and online herbarium sources.

Basin wildrye occurs in 2 cytotypes, tetraploid (2n = 28) and octoploid (2n = 56) (Ogle and others 2012). Crossing cytotypes results in sterile seed or unstable hexaploids (Young and others

2013). Consequently, cytotypes should not be combined when assembling multi-origin seed sources for grow out or on restoration projects. As our interest, in this study, was to evaluate only establishment and short-term persistence, sources (Table 1) were not isolated by cytotype but data were grouped by cytotype for analysis.

In 2013, leaf tissue was collected from CBR populations and processed through a Partec flow cytometer to determine cytotype. Leaves were harvested, placed in resealable plastic bags, and kept on ice while in the field. Leaves were then transferred to a cooler (3 °C [37.4 °F]) until cytotype determination, following methods described by Richardson and others (2012). Leaves from 2 plants per site were processed to determine each population's cytotype. Across the western portion of the species distribution, Culumber and others (2013) found no cases of mixed ploidy within populations of basin wildrye.

When developing genetically appropriate stock seed supplies, leading recommendations suggest pooling seed from at least 5 populations representing in aggregate more than 50 parents (Brown and Marshall 1995; Withrow-Robinson and Johnson 2006; Johnson and others 2010b). Basin wildrye populations were mapped according to seed zone and cytotype to identify populations for each zone/cytotype combination. For this study, we were not able to locate 5 populations for each

TABLE 1

Locations, cytotypes, and generalized provisional seed zone of basin wildrye populations and test sites used in this study.

Basin wildrye populations	Latitude	Longitude	Cytotype	Provisional seed zone	County, State
Bartine Ranch	39.53	-116.34	Octoploid	10–15 Deg. F. / 6–12	Eureka, Nevada
Crescent Valley	40.56	-116.14	Octoploid	10–15 Deg. F. / 6–12	Elko, Nevada
West Elko	40.82	-115.80	Octoploid	10–15 Deg. F. / 6–12	Elko, Nevada
Willow Creek	41.29	-114.84	Octoploid	10–15 Deg. F. / 6–12	Elko, Nevada
Secret Creek	40.87	-115.32	Tetraploid	10–15 Deg. F. / 6–12	Elko, Nevada
Walti Ranch	39.94	-116.57	Tetraploid	10–15 Deg. F. / 6–12	Eureka, Nevada
Wilkins	41.37	-114.80	Tetraploid	10–15 Deg. F. / 6–12	Elko, Nevada
Angel Creek	41.05	-115.02	Tetraploid	15–20 Deg. F. / 3–6	Elko, Nevada
Grasshopper Spring	39.22	-114.69	Tetraploid	15–20 Deg. F. / 3–6	White Pine, Nevada
Pinto Summit	39.45	-114.94	Tetraploid	15–20 Deg. F. / 3–6	Eureka, Nevada
Rosebud	39.01	-114.67	Tetraploid	15–20 Deg. F. / 3–6	White Pine, Nevada
Windfall Canyon	39.46	-115.97	Tetraploid	15–20 Deg. F. / 3–6	Eureka, Nevada
Burge Ranch	41.47	-117.53	Octoploid	15–20 Deg. F. / 6–12	Humboldt, Nevada
Grass Valley Road	40.51	-117.60	Octoploid	15–20 Deg. F. / 6–12	Pershing, Nevada
Melendez Ranch	38.93	-115.17	Octoploid	15–20 Deg. F. / 6–12	White Pine, Nevada
Paradise Valley	41.52	-117.54	Octoploid	15–20 Deg. F. / 6–12	Humboldt, Nevada
Antelope	39.40	-115.47	Tetraploid	15–20 Deg. F. / 6–12	Eureka, Nevada
Eureka	39.51	-115.96	Tetraploid	15–20 Deg. F. / 6–12	Eureka, Nevada
Gleason Creek	39.46	-115.04	Tetraploid	15–20 Deg. F. / 6–12	White Pine, Nevada
King Top Mountain	39.04	-113.62	Tetraploid	15–20 Deg. F. / 6–12	Millard, Utah
Pleasant Valley	40.25	-117.67	Tetraploid	15–20 Deg. F. / 6–12	Pershing, Nevada
Grass Valley	39.47	-117.00	Octoploid	20–25 Deg. F. / 6–12	Lander, Nevada
Eldorado Canyon	40.52	-118.23	Tetraploid	20–25 Deg. F. / 6–12	Pershing, Nevada
Twentyone Mile Hill	39.77	-119.69	Tetraploid	20–25 Deg. F. / 6–12	Washoe, Nevada
West of Water Canyon	40.61	-116.68	Tetraploid	20–25 Deg. F. / 6–12	Lander, Nevada
Test sites	Latitude	Longitude		Provisional seed zone	County, State
Nephi	39.64	-111.87		15–20 Deg. F. / 6–12	Juab, Utah
Fountain Green	39.61	-111.61		10–15 Deg. F. / 6–12	Sanpete, Utah
Spanish Fork	40.07	-111.62		15–20 Deg. F. / 3–6	Utah, Utah
Orovada	41.53	-117.75		20–25 Deg. F. / 6–12	Humboldt, Nevada

GPSZ/cytotype combinations (Table 2), nevertheless all had in excess of 100 parents.

Magnar, Trailhead, and other source-identified populations of basin wildrye have historically been seeded extensively in the Great Basin, especially following fire. To minimize the possibility of collecting planted sources of these materials, we excluded populations that occurred within evident fire perimeters, adjacent to roadsides, and within boundaries of federal restoration projects (Pilliod and Welty 2013), and we visited with landowners or ranch managers about seeding history on private lands.

In 2013 and 2014, basin wildrye seed was harvested from a minimum of 50 individual plants at 25 wildland populations. Seed from each source was maintained separately, cleaned

TABLE 2

Local provisional seed zone for each test site along with the number of basin wildrye populations of each cytotype originating from that provisional seed zone and pooled for comparison. Also shown are the comparisons that were made to non-local sources and cytotypes.

Site	Local source	n	Cytotype	Non-local sources							
Orovada	20–25 Deg. F. / 6–12	3	Tetraploid		15–20 Deg. F. / 6–12	15–20 Deg. F. / 3–6	10–15 Deg. F. / 6–12	Tetra			Trail
		2	Octoploid		15–20 Deg. F. / 6–12		10–15 Deg. F. / 6–12		Cont	Mag	
Nenni	15–20 Deg.	3	Tetraploid	20–25 Deg. F. / 6–12		15–20 Deg. F. / 3–6	10–15 Deg. F. / 6–12	Tetra			Trail
	F. / 6–12	3	Octoploid	20–25 Deg. F. / 6–12			10–15 Deg. F. / 6–12		Cont	Mag	
Snanish Fork	15 20 Deg.	5	Tetraploid	20–25 Deg. F. / 6–12	15–20 Deg. F. / 6–12		10–15 Deg. F. / 6–12	Tetra			Trail
	F. / 3–6	0	Octoploid						Cont	Mag	
Fountain Green	10–15 Deg.	4	Tetraploid	20–25 Deg. F. / 6–12	15–20 Deg. F. / 6–12	15–20 Deg. F. / 3–6		Tetra			Trail
	F. / 6–12	5	Octoploid	20–25 Deg. F. / 6–12	15–20 Deg. F. / 6–12				Cont	Mag	

Notes: Shading indicates data were not available at these sites due to seeding failures. Tetra = Great Basin Tetra; Cont = Continental; Mag = Magnar; Trail = Trailhead.



116 Orovada, Nevada, test plot planted and covered with row cover fabric.

using a Clipper laboratory seed cleaner and Carter Day fractionating aspirator, and viability tested (TZ) at the Utah State Seed Laboratory. Commercial certified basin wildrye sources, Magnar, Trailhead, and Tetra, were acquired from the Utah Division of Wildlife Resources, Great Basin Research Center, seed warehouse.

Test Sites

Test sites were located in each of the 4 GPSZ's with basin wildrye population occurrence rate more than 5%. The Spanish Fork, Utah, location represented zone 15–20 Deg. F. / 3–6 and occurred on a Timpanogos loam soil type (Farmlogs 2018a). Annual precipitation averaged 43 cm (17 in) over the past 10 y (Farmlogs 2018b). The Nephi, Utah, location represented zone 15–20 Deg. F. / 6–12 and occurred on Nephi silt loam soil. It averaged 28 cm (11 in) of precipitation. The Orovada, Nevada, location represented zone 20–25 Deg. F. / 6–12 and occurred on the Snapp-McConnel-Adeliade soil association. It averaged 16.5 cm (6.5 in) of annual precipitation. The Fountain Green, Utah, location represented the 10–15 Deg. F. / 6–12 GPSZ. Soils were classified as a Keigley silty clay loam (Farmlogs 2018a), and annual precipitation averaged 33 cm (13 in) over the past 10 y (Farmlogs 2018b).

In May of 2013 and 2014, plots at each site were disked to incorporate existing vegetation and summer fallowed. In early fall, the sites were harrowed and roller packed to prepare the seedbed for planting. Several weeks prior to planting, plots were treated with the nonselective herbicide glyphosate to remove fall-germinated seedlings. Glyphosate was chosen for its ability to control a broad spectrum of grasses and broadleaf plants while also strongly binding to soil particles, rendering it biologically unavailable with no soil residual activity (Monsanto 2004). Planting occurred during the first half of November in both years. In 2014, plots were mowed in May at a 10 cm (4 in) stubble height to limit weed seed maturation. Plots were not mowed in 2015.

Row Cover Treatment

DeWitt's 51 g/m² (1.5 oz/yd²) N-Sulate (DeWitt Company, Sikeston, Missouri) is a medium weight, permeable, UV-treated fabric designed to offer frost protection to plants, reduce evaporation rate, lengthen harvest time, and/or extend flowering season. Studies report 3 to 5 °C higher soil temperatures under row cover (Harris and others 2015) and up to 6% more moisture (Tilley and others 2009) compared to uncovered soils. Other studies report increased survival of spring emerging seedlings (Tilley and others 2009; Stettler and Whitaker 2012; Shock and others 2013) as well as overwintering perennial plants (Harris and others 2015). Following seeding, beds were enclosed with 3.6 m (12 ft) wide row cover using a tractor-drawn plastic mulch layer. The cover remained in place through early spring.

Study Design

A primary objective of this study was to compare establishment success among populations seeded in a manner similar to a restoration planting. This approach differs from most reciprocal studies, which are established using outplanted nurserygrown stock. Outplants, particularly suitable from a research design perspective, permit evaluation of mature plant performance and longevity, but bypasses the seedling life phase. Establishing the study by direct seeding permits evaluation of this critical establishment phase, upon which all subsequent performance and longevity data rely.

The study was implemented as a 2-factor factorial that included row cover treatment and seed source in a split plot design. We assigned 10 whole plots (blocks) for treatments of row cover (n = 5 blocks) or uncovered controls (n = 5 blocks), and seed sources were randomly assigned in subplots. In 2013 and 2014, at each site, each of 29 seed sources was planted at a density of 68.3 pure live seeds (PLS) per m (20.83 PLS/ft) along a 1.8 m (6 ft) row length. Seed was sown 2 cm (0.8 in) deep with a tractor-drawn custom Hege 1000 series cone seeder, planting through John Deere double disc openers.

Data Collection and Analysis

Year 1 establishment data, recorded as the total number of individual plants occurring in each row, were collected in the fall of each year following seeding. Short-term persistence data (year 3-4), measured as linear centimeters of row occupied by basin wildrye, were recorded in fall 2017, 4 y post-2013 planting and 3 y post-2014 planting. Sources were classified as local if they originated in the same GSPZ as the test site. For analysis, data from individual sources were either evaluated separately or pooled by cytotype or originating GPSZ to address hypotheses of interest. For example, when evaluating establishment differences at Fountain Green between local (10-15 Deg. F. / 6-12) octoploids and non-local (15-20 Deg. F. / 6-12) octoploids, data from 5 populations were grouped to create the local octoploid data set, and data from 3 populations were grouped to create the non-local octoploid data set. Tetraploid and octoploid groupings were analyzed separately because of the genetic constraints of mixing these cytotypes. Data were analyzed using SAS Mixed Procedure (SAS Institute 2004) as a split plot, with cover treatments as whole plots and sources as subplots. Site, year, cover treatment, and source were considered fixed effects whereas block was considered a random effect.

Precipitation data were assembled from Farmlogs by delineating study sites and using the built-in rainfall feature. Farmlogs report National Oceanic and Atmospheric Administration (NOAA) data, which are calculated using radar and ground stations to algorithmically predict the amount of precipitation that falls on a high-resolution (1 km [0.6 mi]) grid of the US.

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RESULTS

Year 1 Establishment

At both the Orovada and Spanish Fork sites in 2013 and 2014, spring seedling emergence was very poor and subsequent mortality was high. By fall, when year 1 data were recorded, too few plants remained to make valid statistical inferences. For the remaining sites, we compared year 1 survival between pooled local GPSZ sources (occurring within the same GSPZ as the test site) and non-local (originating from other GPSZ) tetraploid sources, pooled local and nonlocal octoploid sources, and local tetraploid groups or local octoploid groups versus commercial sources (see Table 2). At Fountain Green, significant differences were observed between sources, treatments, year, and the interaction treatment \times year (Table 3). Magnar established better than pooled local octoploids and pooled local tetraploids (Table 4). No differences occurred between local and non-local pooled GPSZ sources or pooled local GPSZ sources and other commercial sources. Survival in row cover treatments was better in 2014 than 2013 (Table 5), and survival in row cover treatments was better both years than survival in control plantings in 2014 and 2013, which also differed from each other. Contrasting with row cover treatments, survival in control was higher in 2013 than 2014 causing the significant treatment × year interaction (Tables 3 and 5).

For year 1 survival at Nephi, significant differences were observed between treatment, year, and the interaction treatment \times year (Table 3). No detectable differences occurred between sources. Similar to the Fountain Green site, survival in row cover treatments was better in 2014 than 2013, and survival in row cover treatments was better both years than survival in control plantings, which did not differ by year (Table 5). Contrasting with row cover treatments, survival in control was greater, though not significantly so, in 2013 than in 2014 (Table 5), causing the significant treatment \times year interaction (Table 3).

Short-Term Persistence

At Fountain Green there were no differences among sources in short-term persistence (year 3–4), measured as linear centimeters of row occupied by basin wildrye (Table 3). We observed significant differences in short-term persistence between treatment, year, and the treatment × year interaction (Table 3). Short-term persistence was higher in row cover than in control plots both years. Short-term persistence was higher in 2013 than in 2014 row cover plots. Control plots were similar between years. In contrast to row cover treatments, survival in control was greater in 2014 than in 2013, though not significantly so, causing the significant treatment × year interaction (Table 5). At Nephi, differences were observed only between

TABLE 3

Values (F and P) for testing overall effects of year 1 establishment for basin wildrye populations.

Site		Year 1 esta	blishment	Short-term persistence		
	Effect	F statistic	P value	F statistic	P value	
Fountain Green						
	Source	1.76	0.0104	0.918	0.5885	
	Treatment	372.86	<.0001	56.871	<.0001	
	Source × Treatment	1.21	0.2103	4.011	0.9798	
	Year	11.08	0.0009	0.525	0.0458	
	Source × Year	1.15	0.2730	0.480	0.9898	
	Treatment × Year	26.14	<.0001	5.867	0.0158	
	Source × Treatment × Year	0.98	0.4980	0.408	0.9973	
Nephi						
	Source	1.19	0.2376	0.283	1	
	Treatment	422.12	<.0001	2.332	0.127	
	Source × Treatment	1.16	0.2665	92.717	0.999	
	Year	227.62	<.0001	0.354	<.0001	
	Source × Year	0.7	0.872	0.247	1	
	Treatment × Year	238.17	<.0001	0.004	0.948	
	Source × Treatment × Year	0.76	0.8106	0.242	1	

Site	Cytotype	Source comparisons	Estimate	DF	t Value	Pr> t
Fountain Green (10–15 Deg F / 6–12)						
	Octoploids	Local vs 15 – 20 Deg. F. / 6 – 12	0.0330	460	0.02	0.9859
		Local vs 20 – 25 Deg. F. / 6 – 12	0.5250	460	0.24	0.8080
		No sources from 15 – 20 Deg. F. / 3 – 6				
		Local vs Continental	0.2500	460	0.09	0.9296
		Local vs Magnar	-10.1500	460	-3.56	0.0004
		Local vs Tetra	-3.1500	460	-1.11	0.2657
		Local vs Trailhead	-5.0000	460	-1.77	0.0776
	Tetraploids	Local vs 15 – 20 Deg. F. / 3 – 6	-3.5250	460	-1.93	0.0540
		Local vs 20 – 25 Deg. F. / 6 – 12	-2.8375	460	-1.55	0.1206
		Local vs 15 – 20 Deg. F. / 6 – 12	1.1333	460	0.58	0.5656
		Local vs Continental	0.3500	460	0.12	0.9035
		Local vs Magnar	-10.0500	460	-3.48	0.0005
		Local vs Tetra	-3.0500	460	-1.06	0.2910
		Local vs Trailhead	-4.9000	460	-1.7	0.0901
Nephi (15–20 Deg F / 6–12)						
	Octoploids	Local vs 10 – 15 Deg. F. / 6 – 12	-1.3767	460	-1.06	0.2881
		Local vs 20 – 25 Deg. F. / 6 – 12	-0.7417	460	-0.46	0.6469
		No sources for 15 – 20 Deg. F. / 3 – 6				
		Local vs Continental	-2.5167	460	-1.23	0.2194
		Local vs Magnar	-0.9667	460	-0.47	0.6369
		Local vs Tetra	-1.8167	460	-0.89	0.3752
		Local vs Trailhead	-2.5167	460	-1.23	0.2194
	Tetraploids	Local vs 10 – 15 Deg. F. / 3 – 6	2.3625	460	1.89	0.0601
		Local vs 15 – 20 Deg. F. / 3 – 6	-0.3000	460	-0.024	0.8109
		Local vs 20 – 25 Deg. F. / 6 – 12	1.0958	460	0.81	0.4186
		Local vs Continental	-0.0375	460	-0.02	0.9849
		Local vs Magnar	1.5125	460	0.76	0.4457
		Local vs Tetra	0.6625	460	0.33	0.7383
		Local vs Trailhead	-0.0375	460	-0.02	0.9849

Values (t and P) for testing source effects on year 1 establishment of basin wildrye populations.

years (Table 3), with short-term persistence better in 2014 than in 2013 for both treatments (Table 5).

DISCUSSION

Unfortunately, failed plantings at Orovada and Spanish Fork reduced the scope of inference to half of what was intended. The ultimate cause of failure is unknown, although observed emergence was minimal at these sites. We speculate seed may have germinated during fall or winter months and perished from desiccation or frost (Roundy and Madsen 2016), or competition from weedy species may have suppressed emergence (Thomson and others 2017). Basin wildrye is a poor competitor during establishment when growing with competitive species (Ogle and others 2012), and at both the Orovada and the Spanish Fork sites, even following efforts to reduce the seedbank, considerable competition occurred from weedy species. Both locations appear suitable to support this grass: Remnant native stands persist adjacent to the Orovada test site on similar soil and topography and a planted field of basin wildrye persists near the Spanish Fork plot on similar soils. With these sites failing to contribute data to the objective, their singular contribution is merely to provide further observation that basin wildrye is a poor competitor during establishment.

A primary objective of this study was to test whether GPSZs geographically represent relevant partitions of selective gradi-

Values (t and P) for testing Treatment × Year effects on year 1 es-
tablishment and year 3–4 persistence of basin wildrye populations.

Site	Treatmer Year effe		Estimate	Pr> t	
Fountain Green					
Year 1					
	Row cover	2013	17.7034 B ^z	<.0001	
	Row cover	2014	25.7931 A	<.0001	
	Control	2013	4.0966 C	0.0149	
	Control	2014	2.3862 D	0.1551	
Year 3–4					
	Row cover	2013	91.1862 A	<.0001	
	Row cover	2014	69.8689 B	0.0012	
	Control	2013	43.1862 C	0.3692	
	Control	2014	45.2068 C	0.7385	
Nephi					
Year 1					
	Row cover	2013	4.5931 B	<.0001	
	Row cover	2014	24.6828 A	<.0001	
	Control	2013	1.2276 C	0.1058	
	Control	2014	1.0000 C	0.1875	
Year 3–4					
	Row cover	2013	20.1793 B	0.0138	
	Row cover	2014	61.1862 A	<.0001	
	Control	2013	13.3517 B	0.0277	
	Control	2014	54.9172 A	<.0001	

^zLetters indicate significant differences within sites.

ents under which basin wildrye has evolved, and thereby aptly partition populations into adaptive groups. We hypothesized local adaptation expressed by pooled local sources would perform better initially and over time than pooled non-local sources. Additional objectives were to compare local pooled GPSZ sources to the commercially available sources Trailhead, Magnar, Continental, and Great Basin Tetra and to evaluate row cover as a method to improve seedling establishment.

We expected local adaptation would first be expressed by differential establishment rates with local climate benefiting local sources. But, we observed no differences between local and non-local sources, grouped by cytotype and pooled by GPSZ, either initially or 3 to 4 y later. To ensure results from pooled sources were not obscured by poor performing individual sources, we compared pooled local GPSZ sources to local individual sources and again found no source differences. In other words, local individual sources performed similarly to local pooled sources at these 2 sites. Commercial sources, excepting Magnar, also failed to distinguish themselves. Magnar initially established better than local sources at Fountain Green but that difference was no longer evident by year 3-4. The intent of seed zones is to match seed sources to geographic areas where they are well adapted. Over the brief duration of this study, the lack of differences suggests basin wildrye functioned as a habitat generalist, unresponsive to seed zones. Selective gradients, if present or expressed during the study interval, did not exert persistent differential results. It is likely that documenting such selectivity may require more annual replication and longer persistence intervals to capture climatic variability sufficient to cause selective pressure. For basin wildrye, empirical seed zones are now available (Johnson and Vance-Borland 2016)



120 Fountain Green, Utah, test plot following several years of growth.

and are the best available tool to guide source development and deployment.

Row cover treatments dramatically improved year 1 establishment as compared to control treatments. The most modest improvement was 318% at Nephi in 2013 (see Table 5). Good germination was observed in both row cover and control plots; however, the combined benefits of row cover, frost protection, warmer temperatures, and prolonged moisture availability dramatically benefited seedling establishment. These positive benefits are similar to others' findings (Tilley and others 2009; Stettler and Whitaker 2012; Shock and others 2013). At Fountain Green, beneficial effects remained through year 3–4, whereas at Nephi they were no longer statistically better. Using row cover to improve establishment may be suitable for some projects.

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